

## Oxygen depletion in lake hypolimnia

Brian Rippey\* and Catherine McSorley

School of Environmental Sciences, University of Ulster, Coleraine, Northern Ireland, United Kingdom

### Abstract

The Livingstone and Imboden (1996) model for the oxygen depletion rate in lake hypolimnia was evaluated with field results in 32 lakes. The volumetric oxygen consumption rate ( $J_V$ ) was strongly related to lake trophic state, as represented by the growing season mean chlorophyll *a* (Chl *a*) and annual mean total phosphorus concentration, and might reach a maximum of 0.23–0.24 g O<sub>2</sub> m<sup>-3</sup> d<sup>-1</sup> in eutrophic lakes. The areal oxygen consumption rate ( $J_A$ ) was not related to trophic state and was lower than the value estimated by the sediment core method, particularly in eutrophic lakes; the mean was  $0.0816 \pm 0.0150$  g O<sub>2</sub> m<sup>-2</sup> whole sediment d<sup>-1</sup>. We modified the model and it was used to confirm that the field  $J_A$ , the average for the hypolimnion for the deoxygenation period, is influenced by  $J_V$  and the range of sediment area to water volume ratio,  $\alpha(Z)$  ( $Z$  is water depth), in the hypolimnion, as well as the oxygen consumption rate in whole sediment ( $R$ ). It was suggested that this is the reason for the difference between field and sediment core  $J_A$  values. The  $R$  in 10 lakes varied from 7 to 460 g O<sub>2</sub> m<sup>-3</sup> whole sediment d<sup>-1</sup>. Overall, the Livingstone and Imboden model with constant  $J_A$  and with a  $J_V$  value derived from an empirical relationship with either measure of trophic state was considered to be a suitable condensed model for oxygen depletion in lake hypolimnia.

### Introduction

The depletion of oxygen in lake water, particularly in the hypolimnion, was one of the characteristics of lakes studied by the early limnologists. For example, Strom (1931) introduced the areal hypolimnetic oxygen deficit (AHOD) to summarize the behavior of oxygen in hypolimnia, and Hutchinson (1938) and Mortimer (1942) used it to classify lakes. Hutchinson (1957) has summarized this early research, after which the focus was to relate AHOD to measures of lake trophic state and to understand the influence of lake morphometry and other properties. The final result of this effort was the statistical model of Cornett and Rigler (1980), which allows AHOD to be predicted from lake water annual mean total phosphorus concentration, lake mean depth, and water temperature in the hypolimnion, although prediction by a nutrient loading model (Welch and Perkins 1979; Vollenweider and Janus 1982) should not be forgotten.

These AHOD models cannot simulate oxygen concentration–depth profiles, except in a general way; the mean oxygen concentration in a hypolimnion of mean depth  $Z$  after a period of days ( $t$ ) would be the initial concentration–(AHOD/ $Z$ )/ $t$ . Cornett (1989), however, did develop a statistical model to predict the oxygen depletion rate in strata in the hypolimnion that used stratum water temperature, stratum water volume to sediment surface area, and phosphorus accumulation rate in the sediment, and it can be used to calculate the oxygen concentration–depth profiles at any time during stratification. Molot et al. (1992) also developed an empirical model to predict the oxygen concentration–depth profile at the end of the summer, and Clarke et al. (2002) have evaluated its wider application. The model uses stratum water volume to sediment surface area ratio, initial oxygen concentration and either the growing season mean epilimnetic total phosphorus concentration or

concentration in spring. During this period, Nurnberg (1995) introduced a new concept, the Anoxic Factor, to summarize the behavior of oxygen in lakes and developed a statistical model to predict its value.

Livingstone and Imboden (1996) developed a deductive model to describe the depletion rate of oxygen in lake hypolimnia. It separates the consumption of oxygen in water from consumption in the sediment to describe the influence of lake morphometry on the oxygen depletion rate. Morphometry was characterized by the sediment area to water volume ratio in the water strata, and the values for the consumption rate in water and sediment apply spatially over the hypolimnion and to the period of deoxygenation, which is seen as an additional advantage of the model (Pace and Prairie 2005). It is almost identical to that of Stefan and Fang (1994), in that they differ only in the degree of aggregation of model parameters. Pace and Prairie (2005) have recently synthesized what is known about respiration in lakes, integrated the results, and developed models to estimate the respiration of organic carbon in the global lake population.

Our motivation for returning to this topic was the European Union's Water Framework Directive (Directive 2000/60/EC), which requires that the quality of the structure and functioning of aquatic ecosystems associated with surface waters is used to establish the ecological quality of surface waters. Although biological properties are the main focus, oxygenation conditions are described as supporting them, and they can also be considered an important structural characteristic of lakes, with oxygen depletion an important process. Whether the hypolimnion is oxygenated, deoxygenated, or intermediate is part of the structure of a lake, in that this affects profundal macroinvertebrates (Dinsmore and Prepas 1997), fish habitat (Evans et al. 1996), and the internal loading and recycling of phosphorus (Nurnberg 1984) and iron and manganese (Davison 1993); the rate of oxygen depletion has already been used to classify lakes (Hutchinson 1957).

\* Corresponding author: bhrt.rippy@ulster.ac.uk

We chose to investigate the Livingstone and Imboden (1996) model further. Those researchers encouraged it, and perhaps our models of the important physicochemical and biological properties of lakes need to be made more deductive or, at least, less empirical if climate change alters the interrelationships between lake properties. Also, perhaps the model has not been fully evaluated because it has only been used to model change in the oxygen profile during the eutrophication of one lake. Therefore, our aim in the work reported here was to evaluate the Livingstone and Imboden model for the oxygen depletion rate in hypolimnia; for this, we collected results from five Irish lakes and collated results for a further 28 from the literature.

## Methods

The five Irish lakes used in this investigation are Lough Alina, Ballydugan Lough, Lough Macrory, Heron Lough, and White Lough. They are small lakes in the deglaciated landscape of Northern Ireland, with areas of 7, 12, 21, 5, and  $6 \times 10^4$  m<sup>2</sup>, respectively, and they were visited weekly between 06 April and 07 June 2005. The mean and maximum water depths are 5.5 and 14.0 m, respectively, in Alina, 4.9 and 12.0 m in Ballydugan, 2.8 and 10.5 m in Macrory, 3.9 and 10.5 m in Heron, and 6.2 and 10.7 m in White Lough. The thermal stratification in Lough Macrory was too weak to estimate the oxygen depletion rate, but the areal consumption rate was still estimated from sediment cores. The lakes were chosen so that they covered a range of trophic states, as represented by the annual mean total phosphorus concentration, which was 11, 20, 38, 40, and 110  $\mu\text{g P L}^{-1}$  in Macrory, Alina, White, Ballydugan, and Heron, respectively.

Water samples from the mixed layer were taken with a Ruttner sampler, and a composite was produced. The concentration of chlorophyll *a* (Chl *a*) was determined by hot methanol extraction (Marker et al. 1980). The temperature and oxygen concentration depth profiles were recorded at a central site with a YSI MP556 Multi-Probe System with barometer and a YSI 5562 oxygen–temperature probe with 20 m of cable. The probe was calibrated in the field at each lake with water-saturated air and the barometric pressure.

The temperature results were used to estimate the eddy diffusion coefficient of heat for the main period of deoxygenation according to the flux-gradient method (Jassby and Powell 1975).

The Livingstone and Imboden (1996) model was used with the oxygen results for four of the Irish lakes to estimate values for the volumetric oxygen consumption rate and areal oxygen consumption rate. This model of the depletion of oxygen in the hypolimnion of a lake is

$$J(Z) = J_V + J_A \alpha(Z) \quad (1)$$

where  $J(Z)$  is the oxygen depletion rate ( $\text{g O}_2 \text{ m}^{-3} \text{ water d}^{-1}$ ) at water depth  $Z$  (m of water),  $J_V$  is the volumetric oxygen consumption rate ( $\text{g O}_2 \text{ m}^{-3} \text{ water d}^{-1}$ ),  $J_A$  is the areal oxygen consumption rate ( $\text{g O}_2 \text{ m}^{-2} \text{ whole sediment d}^{-1}$ ), and  $\alpha(Z)$  is the sediment surface area to water volume ratio ( $\text{m}^2 \text{ sediment m}^{-3} \text{ water}$ ) at water depth  $Z$ .  $J_V$  and  $J_A$

are assumed to be constant with water depth and time and the vertical diffusion of oxygen is assumed to be negligible. The slope of the linear regression of  $J(Z)$  against  $\alpha(Z)$  is, therefore, an estimate of  $J_A$  and the intercept an estimate of  $J_V$ . The consumption of oxygen in the sediment as a proportion of the total consumption in the water stratum at water depth  $Z$  is  $(\alpha(Z)J_A)/J(Z)$ .

Estimates of  $J_V$  and  $J_A$  in other lakes were also made with the use of field results collated from the literature. The Web Appendix ([http://www.aslo.org/lo/toc/vol\\_54/issue\\_3/0905a.pdf](http://www.aslo.org/lo/toc/vol_54/issue_3/0905a.pdf)) gives the sources of the results and how  $J(Z)$  and  $\alpha(Z)$  were derived. Results from 28 additional lakes, mostly from Canada, were available.

Between 3 and 14 data points were included in the regressions of  $J(Z)$  against  $\alpha(Z)$  (mean = 7.7, median = 8), and so the statistical power of some regressions to establish small  $J_A$  values was low. For example, for  $\alpha = 0.05$ , between 30 and 300 data points are needed for a test to establish a product moment correlation coefficient between 0.30 and 0.10, respectively, with a power of 0.50; for  $\alpha = 0.10$ , 19–180 data points are needed (Cohen 1988). As a result, we used an  $\alpha$  significance criterion of  $>0.05$  in some regressions;  $\alpha$  was  $<0.05$  in 17 lakes,  $>0.05$  but  $<0.29$  in seven, and  $>0.29$  in the remainder. When the regression between  $J(Z)$  and  $\alpha(Z)$  was not considered significant, and so the value of  $J_A$  was indistinguishable from zero, the mean of the  $J(Z)$  values was used as an estimate of  $J_V$ .

$J_A$  in the Irish lakes was also estimated from sediment cores in the laboratory according to the method of Sweerts et al. (1989). The sediment cores were collected with a modified Kajak–Brinkhurst gravity corer, and three replicate cores were taken from sediment in the upper part of the hypolimnion—three from the middle and three from the deepest part—between 21 July and 23 August 2005. They were incubated at the average temperature of the hypolimnetic water (9.1°C in Alina, 8.7°C in Ballydugan and Heron, 10.2°C in Macrory, and 6.5°C in White) and aerated overnight. A stirring rate of 10 rpm was used to produce a diffusive boundary layer thickness of around 1 mm, which is considered typical of the conditions in the hypolimnion of lakes, and the oxygen concentration was measured with a HACH HQ10 oxygen–temperature probe.  $J_A$  was not correlated with water depth in any lake;  $p$  varied between 0.203 and 0.924 (average = 0.513, median = 0.566). The coefficient of variation of  $J_A$  ( $n = 9$ ) for each lake was between 10.1% and 14.4%, and the standard error was between 0.019 and 0.036  $\text{g O}_2 \text{ m}^{-2} \text{ d}^{-1}$  (mean = 0.028, median = 0.027). Graneli (1977) also found low spatial variability of  $J_A$  measured with sediment cores in three lakes; the standard error was between 3.1% and 10.5% of the mean value.

The Livingstone and Imboden (1996) model was modified as follows.  $J_A$  has been modeled by Bouldin (1968), and his steady-state model based on a zero-order reaction rate of oxygen consumption (Eqs. 2–5) has been widely used.

$$J_A = [2R\phi C(0)D_S]^{0.5} \quad (2)$$

$$L = [2\phi C(0)D_S/R]^{0.5} \quad (3)$$

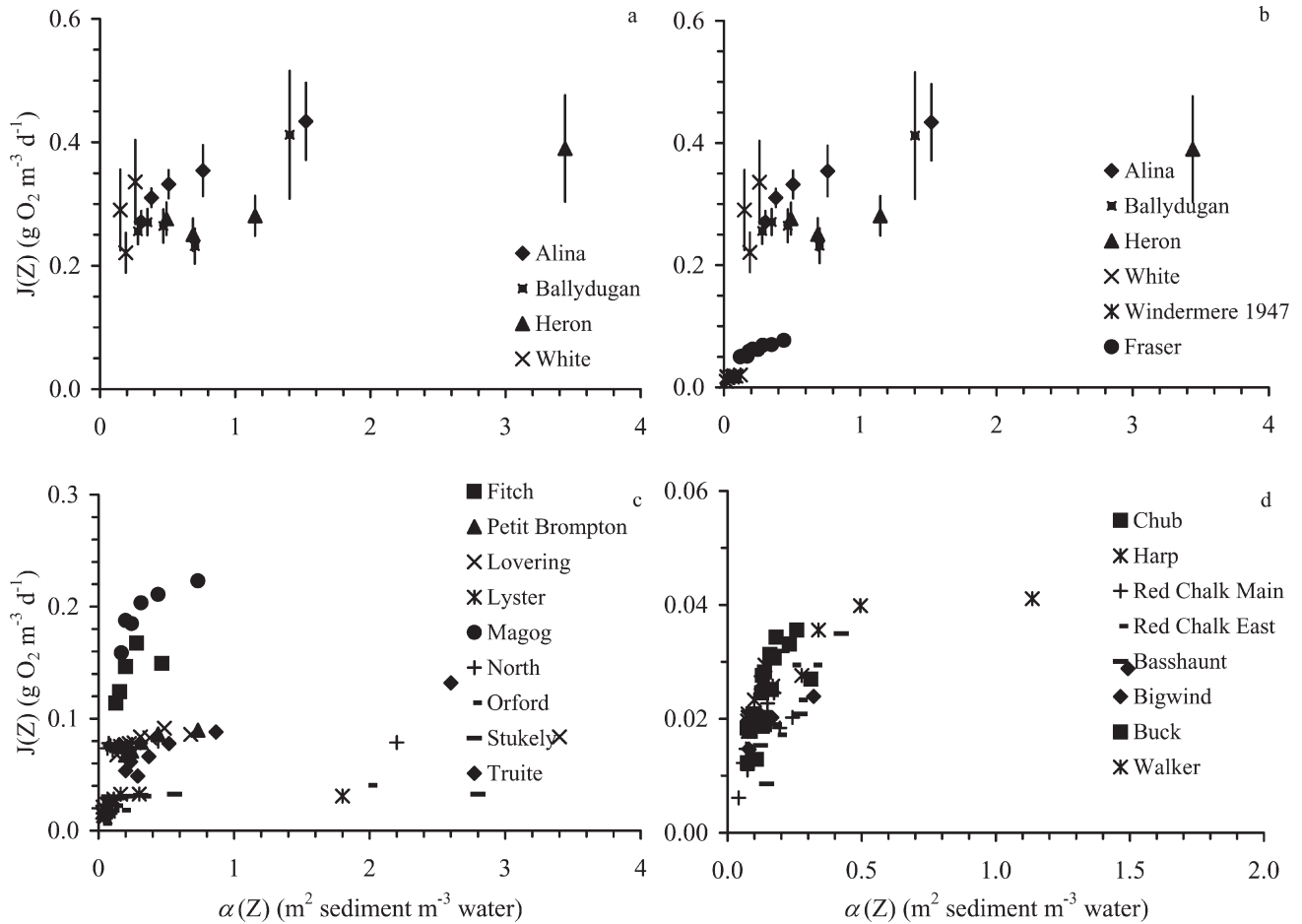


Fig. 1. The variation of oxygen depletion rate,  $J(Z)$ , at water depth  $Z$  with the sediment area to water volume ratio,  $\alpha(Z)$ , in strata in the hypolimnion of lakes. (a) Four Irish lakes, Alina, Ballydugan, Heron, and White. (b) Four Irish lakes and Lake Windermere North and Lac Fraser, Quebec. (c) Nine Quebec lakes. (d) Eight central Ontario lakes. Where shown, the error is standard error.

$$J_A = RL \tag{4}$$

$$C(z) = (R/2\phi D_S)z^2 - [2RC(0)/\phi D_S]^{0.5}z + C(0) \tag{5}$$

$C(z)$  is the oxygen concentration in pore water ( $\text{g O}_2 \text{ m}^{-3}$  water) at depth  $z$  in the sediment (m whole sediment),  $\phi$  is the porosity of the sediment ( $\text{m}^3 \text{ water m}^{-3}$  whole sediment),  $R$  is the zero-order rate of oxygen consumption in the whole sediment ( $\text{g O}_2 \text{ m}^{-3}$  whole sediment  $\text{d}^{-1}$ ),  $C(0)$  is the oxygen concentration at the sediment–water interface ( $\text{g O}_2 \text{ m}^{-3}$  water),  $D_S$  is the diffusion coefficient of oxygen in the whole sediment ( $\text{m}^2$  whole sediment  $\text{d}^{-1}$ ), and  $L$  is the depth of oxygen penetration in the sediment (m whole sediment). Steady-state conditions with  $\phi$ ,  $R$ , and  $D_S$  constant with depth are assumed. Hall et al. (1989) then improved the Bouldin model by including a diffusive boundary layer above the sediment–water interface (Eq. 6).

$$J_A = [(R\delta_D D_S)/D] \left( 1 - \left\{ [2D^2\phi C_w] / [(RD_S\delta_D^2) + 1] \right\}^{0.5} \right) \tag{6}$$

$C_w$  is the oxygen concentration in the water stratum ( $\text{g O}_2$

$\text{m}^{-3}$  water) above the diffusive boundary layer,  $\delta_D$  is the thickness of the diffusive boundary layer (m water), and  $D$  is the diffusion coefficient of oxygen in water ( $\text{m}^2 \text{ water d}^{-1}$ ). Rearranging Eq. 2, the oxygen concentration at the sediment surface at the bottom of the diffusive boundary layer is as shown in Eq. 7.

$$C(0) = J_A^2 / (2R\phi D_S) \tag{7}$$

Values for  $D$  were taken from Broecker and Peng (1974),  $D_S$  was from Maerki et al. (2004), and a temperature of  $10^\circ\text{C}$  and sediment porosity of 0.98 were used;  $D$  was  $0.000151 \text{ m}^2 \text{ water d}^{-1}$ , and  $D_S$  was  $0.000136 \text{ m}^2$  whole sediment  $\text{d}^{-1}$ . A time step of 1 d and an initial value for  $C_w$  of  $10 \text{ g O}_2 \text{ m}^{-3}$  were used.  $\alpha(Z)$  was taken to be in the range  $0.2\text{--}1.0 \text{ m}^2$  whole sediment  $\text{m}^{-3}$  water, with  $\delta_D = 1\text{--}5 \text{ mm}$  and  $J_v \leq 0.3 \text{ g O}_2 \text{ m}^{-3} \text{ d}^{-1}$ .

### Results

The oxygen depletion rate in the hypolimnion,  $J(Z)$ , did not vary greatly with sediment area to water volume ratio,  $\alpha(Z)$ , in the four Irish lakes (Fig. 1a). Only in Lough Alina is the  $J(Z)\text{--}\alpha(Z)$  regression significant at  $p < 0.05$ . The

Table 1. The volumetric oxygen consumption rate ( $J_V$ ) and areal oxygen consumption rate ( $J_A$ ) estimated by the Livingstone and Imboden (1996) model (Eq. 1) with field results in 32 lakes. The standard error (SE) of the value, number of points ( $n$ ), maximum value of sediment surface area to water volume ratio [Max.  $\alpha(Z)$ ], and coefficient of determination ( $R^2$ ) of the regressions are also given. Where the  $J_A$  value was not significantly different from zero at  $p > 0.05$ , but a value is recorded, the  $p$  value is given in parentheses. Where no  $J_A$  value is recorded, the  $p$  value was  $>0.28$ , and where the regression was significant, but  $J_V = 0$ , the  $p$  value for  $J_V$  is given in parentheses. The maximum  $\alpha(Z)$  value shown in parentheses was ignored.

Lake	$J_V$ (g O <sub>2</sub> m <sup>-3</sup> yr <sup>-1</sup> )	SE	$J_A$ (g O <sub>2</sub> m <sup>-2</sup> yr <sup>-1</sup> )	SE	$n$	Max. $\alpha(Z)$	$R^2$
Alina	0.257	0.0139	0.119	0.0169	5	1.52	0.94
Ballydugan	0.200	0.0352	0.135 (0.062)	0.0463	5	1.40	0.74
Heron	0.236	0.0140	0.0442 (0.18)	0.00752	4	3.44	0.95
White	0.282	0.0334			3	0.26	
Windermere North	0.0135	0.00132	0.0592	0.0214	9	0.12	0.52
Fraser	0.0407	0.00246	0.0866	0.00957	9	0.44	0.92
Magog	0.162	0.0110	0.0923	0.0275	6	0.73	0.74
Fitch	0.116	0.0194	0.0977 (0.26)	0.0706	5	0.47	0.39
Truite	0.0523	0.00431	0.0319	0.00401	7	2.60	0.93
Petit Brompton	0.0674	0.00343	0.0324	0.00916	7	0.73	0.71
Lovering	0.0700	0.00309	0.0318	0.00869	9	0.68 (3.40)	0.66
Stukely	0.0279	0.00064	0.00930	0.00262	9	0.56 (2.80)	0.65
Lyster	0.0145	0.00177	0.0746	0.0161	13	0.30 (1.80)	0.66
Orford	0.00905	0.00123	0.0672	0.00962	12	0.33 (2.00)	0.83
Brompton	0.0349	0.00104			13	3.80	
Central Memphremagog	0.0171	0.00215			14	1.70	
Massawippi	0.0191	0.00216			12	1.70	
North Memphremagog	0.0731	0.00142	0.0125 (0.12)	0.00701	8	0.44 (2.00)	0.35
Blue Chalk	0.0317	0.00745			3	0.35	
Chub	0.0160	0.00421	0.0499 (0.10)	0.0272	10	0.31	0.30
Crosson	0 (0.63)		0.227	0.0660	8	0.19	0.66
Dickie	0.0417	0.00860			3	0.54	
Harp	0.00977	0.00271	0.119	0.0257	12	0.17	0.68
Plastic	0.0335	0.0113			3	1.11	
Red Chalk Main	0.00877	0.00244	0.0636	0.0185	10	0.24	0.60
Red Chalk East	0.00390	0.0129	0.0811 (0.24)	0.0492	4	0.33	0.58
Basshaunt	0.0105	0.00470	0.0566	0.0220	9	0.42	0.48
Bigwind	0.0181	0.00062	0.00763	0.00138	12	1.49	0.75
Buck	0.00798	0.00370	0.114	0.0219	8	0.26	0.82
Gullfeather	0.0403	0.00337			3	0.50	
Solitaire	0 (0.10)		0.334	0.0498	6	0.12 (0.27)	0.92
Walker	0.0298	0.00502	0.0111 (0.28)	0.00765	4	1.14	0.51

slope is a field estimate of areal oxygen depletion rate ( $J_A$ ), and the intercept is a field estimate of volumetric oxygen depletion rate ( $J_V$ ). However, if a less severe criterion is used, then the results for Ballydugan Lough ( $p = 0.062$ ) and Heron Lough ( $p = 0.18$ ) also give field estimates of  $J_A$  and  $J_V$  for these lakes (Table 1). The reason for using  $p > 0.05$  with a relatively small number of data points in the regression was described in *Methods*. The regression is not significant ( $p = 0.75$ ) in White Lough, and the small range of  $\alpha(Z)$  (maximum =  $0.26 \text{ m}^2 \text{ m}^{-3}$ ; Table 1) and very small number of data points ( $n = 3$ ) make estimating what must be a low  $J_A$  value difficult;  $J_V$  was calculated as the average of the  $J(Z)$  values (Table 1).

The linear relationship between  $J(Z)$  and  $\alpha(Z)$  in the hypolimnion of Windermere North Basin in 1947 and in Lac Fraser (Quebec, Canada) is shown in Fig. 1b, and the  $J_V$  and  $J_A$  values derived from the regressions are given in Table 1. With the lakes in Quebec there was no sensible relationship between  $J(Z)$  and  $\alpha(Z)$  in Central Memphremagog and Massawippi, so the mean  $J(Z)$  was used as an estimate of  $J_V$  (Table 1). Linear relationships between  $J(Z)$

and  $\alpha(Z)$  in the other lakes (Fig. 1c) were used to estimate field  $J_A$  and  $J_V$  values, in two cases (Fitch and North Memphremagog) with  $0.27 > p > 0.05$  (Table 1). With five of these lakes, the data point with the highest  $\alpha(Z)$  value was ignored ( $3.40 \text{ m}^2 \text{ m}^{-3}$  in Lovering,  $2.80$  in Stukely,  $1.80$  in Lyster,  $2.00$  in Orford, and  $2.20$  in North Memphremagog) because the relationship was strongly linear with the remainder of the results (Fig. 1c). A reason for this behavior is proposed in *Discussion*. The linear relationships in the central Ontario (Canada) lakes were used to derive field  $J_V$  and  $J_A$  values, in three cases (Chub, Red Chalk East, and Walker) with  $0.05 < p < 0.29$  (Table 1). The data point with the highest  $\alpha(Z)$  value in Solitaire was ignored.

Overall,  $J(Z) - \alpha(Z)$  results for 32 lakes were collected or collated from the literature, and 32  $J_V$  and 24  $J_A$  values were derived (Table 1). The range of  $J(Z)$  is large,  $\leq 0.4 \text{ g O}_2 \text{ m}^{-3} \text{ d}^{-1}$ , and the range of the  $\alpha(Z)$  variable varied from  $0.08$  to  $3.80 \text{ m}^2 \text{ m}^{-3}$ , with a typical value of  $0.50 \text{ m}^2 \text{ m}^{-3}$  (Fig. 1).

$J_A$  was also measured in the five Irish lakes by the sediment core method. Because no relationship between  $J_A$

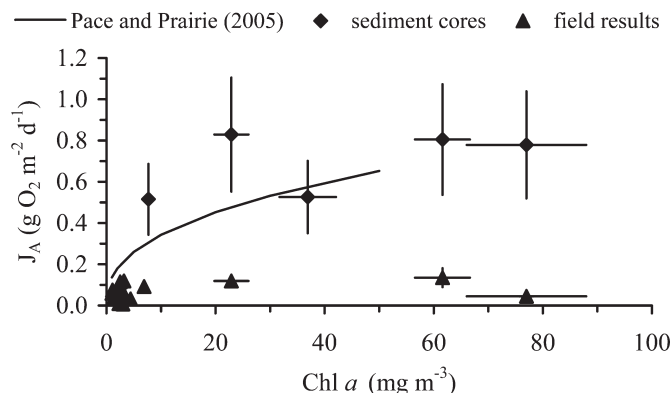


Fig. 2. The variation of areal oxygen consumption rate ( $J_A$ ) with mean growing season lake water chlorophyll *a* (Chl *a*) concentration. The empirical relationship developed by Pace and Prairie (2005) for sediment cores, the value measured from sediment cores in the five Irish lakes, and the value estimated by the Livingstone and Imboden (1996) model (Eq. 1) with field results in four Irish lakes are given. Where shown, the error is standard error.

and water depth appeared in any of these lakes, the mean value for each lake was calculated. It increases with mean growing season Chl *a* concentration, and this behavior agrees with the empirical relationship derived with sediment cores by Pace and Prairie (2005) (Fig. 2). The field  $J_A$  values derived from the Livingstone and Imboden (1996) model (Table 1) are also included in Fig. 2. If the empirical relationship is used to indicate the  $J_A$  value that would be estimated from sediment cores, as was found to be reasonable with the five Irish lakes, then Fig. 2 shows that all but two of the 22 field-derived  $J_A$  values are less than the sediment core values; the sediment core value is greater by between  $-0.16$  and  $0.73$  g O<sub>2</sub> m<sup>-2</sup> d<sup>-1</sup> (mean = 0.19, median = 0.12). The largest difference is found in the three Irish lakes because the field  $J_A$  value does not increase with Chl *a* concentration once the value is  $>5$ – $10$  mg m<sup>-3</sup>. Although the sediment core method attempts to reproduce the hydrodynamic conditions at the sediment–water interface in lakes, as was the case with the method used in the Irish lakes (Sweerts et al. 1989), the difference between the field and sediment core values appears to be too great, particularly in eutrophic lakes, to be caused by the somewhat artificial nature of sediment cores.

One of the assumptions of the Livingstone and Imboden (1996) model that was used to derive the field  $J_A$  and  $J_V$  values is that vertical diffusion of oxygen is negligible. The effect of this on the observed values of  $J(Z)$ , and so on  $J_V$  and  $J_A$  in the Irish lakes, was established by correcting the observed  $J(Z)$  for diffusion. This was completed by using the eddy diffusion coefficients of heat, estimated for each stratum for the main period of deoxygenation from late April to early May to early June 2005. The mean coefficient was  $0.118 \pm 0.022$  m<sup>2</sup> d<sup>-1</sup> in Alina ( $n = 5$ ),  $0.0604 \pm 0.0036$  m<sup>2</sup> d<sup>-1</sup> in Ballydugan ( $n = 5$ ),  $0.0158 \pm 0.0053$  m<sup>2</sup> d<sup>-1</sup> in Heron ( $n = 4$ ), and  $0.056 \pm 0.013$  m<sup>2</sup> d<sup>-1</sup> in White ( $n = 4$ ). The eddy diffusion coefficient of heat was applied to mass and so, because the coefficient for mass can be less than for heat in strongly

thermally stratified layers of water (Quay et al. 1980), the diffusion of oxygen should not be larger than the values based on the diffusion coefficient of heat. Ignoring heat loss to the sediment should also not lead to much of an underestimate of the eddy diffusion coefficient.

The observed vertical oxygen concentration gradient in the strata of a lake, along with areas of the upper and lower boundaries of the strata and the eddy diffusion coefficient, were used to calculate the differences between the masses of oxygen entering and leaving the strata. The differences were then converted to changes in concentration by using the volumes of the strata. There was no relationship between the strata and water depth, so the average net vertical transport of oxygen was calculated for each lake:  $-0.091 \pm 0.024$  g O<sub>2</sub> m<sup>-3</sup> d<sup>-1</sup> in Alina,  $-0.047 \pm 0.0085$  in Ballydugan,  $-0.019 \pm 0.015$  in Heron, and  $-0.001 \pm 0.011$  in White; a negative net vertical transport indicates that transport is toward the sediment surface. These values are small compared with the mean  $J(Z)$ ,  $-0.317 \pm 0.018$  g O<sub>2</sub> m<sup>-3</sup> d<sup>-1</sup> in Alina,  $-0.2557 \pm 0.0085$  in Ballydugan,  $-0.235 \pm 0.035$  in Heron, and  $-0.249 \pm 0.041$  in White (average = 13%; range, 0.22–29%). If  $J(Z)$ , corrected for the net effect of diffusion, is used instead of the observed value, then only small changes in the values of  $J_V$  and  $J_A$  are derived by Eq. 1. The estimates of  $J_V$  are not statistically different, except at Alina, and the mean difference between corrected and observed is  $-0.024$  g O<sub>2</sub> m<sup>-3</sup> d<sup>-1</sup> (range,  $-0.001$  to  $-0.043$  g O<sub>2</sub> m<sup>-3</sup> d<sup>-1</sup>). The mean difference between corrected and observed  $J_A$  values for the three lakes is  $-0.061$  g O<sub>2</sub> m<sup>-2</sup> d<sup>-1</sup> (range,  $-0.033$  to  $-0.079$ ), and the estimates of  $J_A$  made with the observed and corrected  $J(Z)$  values are statistically different, except at Ballydugan.

This analysis of the effect of vertical transport on the estimates of  $J_V$  and  $J_A$  from Eq. 1 in these four eutrophic lakes shows that it alters the values only a little; it changes  $J_V$  by an average of  $-0.024$  g O<sub>2</sub> m<sup>-2</sup> d<sup>-1</sup>, which is small compared with the range of observed values of  $-0.200$  to  $-0.282$  (Table 1). The effect on  $J_A$  is larger: the average change is  $-0.061$  g O<sub>2</sub> m<sup>-2</sup> d<sup>-1</sup> and the observed values are  $-0.0442$ ,  $-0.119$ , and  $-0.135$  for Heron, Ballydugan, and Alina, respectively (Table 1); only in Heron is the effect important, but this is because of the low  $J_A$  value in this lake. However, because  $J_V$  and  $J_A$  were also estimated from the literature, this assumption needs to be assessed further. Although the effect of vertical transport on AHOD has been widely assessed, only Cornett and Rigler (1987) considered the net effect of transport in and out of strata on the depletion rate in the strata; the assessment was completed mainly from oligotrophic lakes, and these are included in Table 1. They found that transport into and out of strata were highly correlated and that net vertical transport was much smaller than the total flux of oxygen. Of the 160 strata in 12 mainly oligotrophic lakes, net transport into these strata was  $>20\%$  of  $J(Z)$  in only 26% of the strata. Their assessment does indicate that the  $J_V$  and  $J_A$  values estimated from field results should not be affected much by assuming that net transport of oxygen is negligible in oligotrophic lakes. Overall, the influence of diffusion is so small in oligotrophic and eutrophic lakes

that this assumption does not reduce the advantage of estimating  $J_V$  and  $J_A$  from field results. The true  $J_A$  values might be a little greater than those estimated by Eq. 1.

## Discussion

**Field values of  $J_V$  and  $J_A$** —The volumetric oxygen consumption rate,  $J_V$ , and areal oxygen consumption rate,  $J_A$ , were derived from the Livingstone and Imboden (1996) model in 32 lakes (Table 1), and these values, along with the sediment area to water volume ratios,  $\alpha(Z)$ , in the hypolimnion, do reproduce the generally linear variation of oxygen depletion rate,  $J(Z)$ , with  $\alpha(Z)$  (Fig. 1). If  $J_A$  is taken to be constant during the deoxygenation period, then the applicability of this model to reproducing field observations is given by the value of the coefficient of determination of the regression (mean = 0.68; median = 0.67; range, 0.30–0.95). Some of the lower values could be caused by the smaller  $\alpha(Z)$  range in those lakes, for the linear regression between coefficient of determination and maximum  $\alpha(Z)$  is significant ( $p = 0.020$ , 23 df). An examination of Fig. 1 does indicate that any departure from linearity in a lake occurs at the upper end of the  $\alpha(Z)$  range; a reason for this is suggested below.

No relationship exists between the field  $J_A$  and  $J_V$  values in the lakes (Fig. 3a); the linear regression is not significant ( $p = 0.85$ , 23 df). This finding is not influenced by the groups of lakes that were used; in particular, the seven lakes at the bottom left of Fig. 3a and the 10 at the upper left are not distinguished by the origin of the results, mainly the 12 lakes in Quebec and 14 in central Ontario.

**Volumetric oxygen consumption rate**— $J_V$  is related to two measures of lake trophic state: the growing season mean Chl *a* concentration (Fig. 3b) and the annual mean total phosphorus concentration (Fig. 3c). The relationship is more precise with Chl *a*, with a suggestion that  $J_V$  reaches a maximum value in eutrophic lakes. Although only the four Irish lakes are considered, this value is  $0.244 \pm 0.0174 \text{ g O}_2 \text{ m}^{-3} \text{ d}^{-1}$  (Table 1). If the relationship with total phosphorus is used, then five values can be employed, the four Irish lakes and Magog, and the maximum value is  $0.227 \pm 0.0212 \text{ g O}_2 \text{ m}^{-3} \text{ d}^{-1}$  (Table 1).

The literature supports the relationship between  $J_V$  and trophic state described by Fig. 3b and c. To use the results in the literature, however, it must be assumed that  $J_V$  is approximately the mean value of  $J(Z)$  for the hypolimnion, which is true in many, but not all, lakes in which  $J(Z)$  does not vary much with  $\alpha(Z)$  (Fig. 1), and that the consumption of oxygen in the sediment,  $\alpha(Z)J_A$ , is small compared with total consumption [mean  $J(Z)$ ]. The mean  $J(Z)$  would be greater than  $J_V$  if consumption in the sediment is not small compared with total consumption.  $J(60 \text{ m})$  in Lake Rotoiti increased from 0.024 to 0.071  $\text{g O}_2 \text{ m}^{-3} \text{ d}^{-1}$  from 1957 to 1984 (see fig. 5 in Vincent et al. 1984), and because the mean Chl *a* concentration was 6.7  $\text{mg m}^{-3}$  during the 1967–1970 period and 9.6  $\text{mg m}^{-3}$  during 1981–1982, these results (Fig. 3d) are consistent with the  $J_V$ –Chl *a* relationship in Fig. 3b. The relationship between the volume-weighted average  $J(Z)$  value in the hypolimnion of five

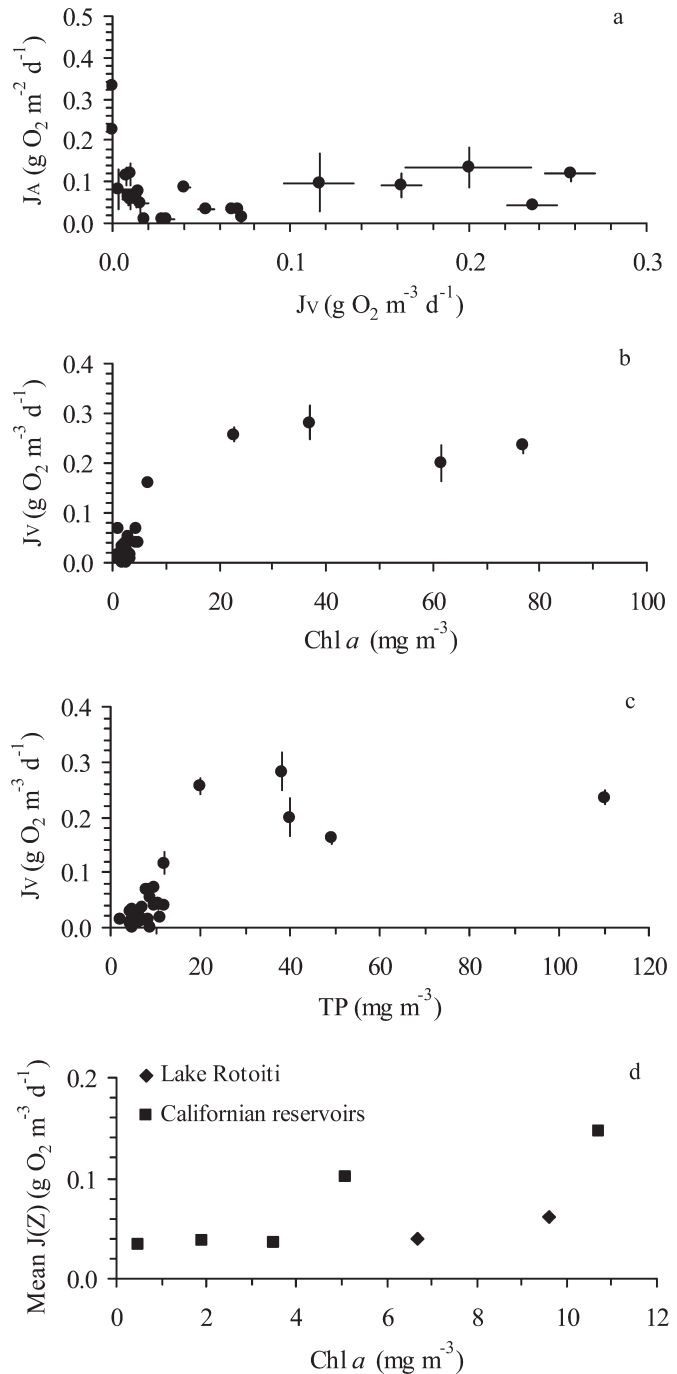


Fig. 3. The variation of (a) areal oxygen consumption rate,  $J_A$ , with volumetric oxygen consumption rate,  $J_V$ , in the hypolimnion of 22 lakes; (b)  $J_V$  with growing season mean lake water chlorophyll *a* (Chl *a*) concentration in 29 lakes; (c)  $J_V$  with annual mean lake water total phosphorus concentration (TP) in 32 lakes; (d) mean oxygen depletion rate in the hypolimnion,  $J(Z)$ , at water depth  $Z$  with Chl *a* in Lake Rotoiti and six reservoirs in California. Where shown, the error is standard error.

reservoirs in California and the Chl *a* concentration (see table 1 in Beutel 2003; Fig. 3d) is also consistent with the relationship in Fig. 3b. Mathias and Barcia (1980) found that, although the volume-weighted average  $J(Z)$  value in the hypolimnion is related to  $\alpha(Z)$ , the average  $J(Z)$  is

higher in eutrophic lakes; lakes with Chl *a* > 5 mg m<sup>-3</sup> and  $\alpha(Z)$  values of 0.6–0.8 m<sup>2</sup> m<sup>-3</sup> have an average  $J(Z)$  value of  $\leq 0.22$  g O<sub>2</sub> m<sup>-3</sup> d<sup>-1</sup>. Finally, the volume-weighted hypolimnetic oxygen concentration at the end of the summer can be used to estimate the mean  $J(Z)$  in 80 lakes in south-central Ontario (Quinlan et al. 2005). Because spring overturn occurs between the last week of April to mid-May and autumn overturn is around 1 September (Molot et al. 1992), the time period available for deoxygenation is 120 d. If the oxygen concentration at spring overturn is 8 g O<sub>2</sub> m<sup>-3</sup>, the average of 14 lakes in central Ontario (Molot et al. 1992), then the mean  $J(Z)$  for the 80 lakes is  $\leq 0.067$  g O<sub>2</sub> m<sup>-3</sup> d<sup>-1</sup>. The maximum value could be higher than this because deoxygenation could have occurred before September in the eutrophic lakes, but these values are generally consistent with the relationship in Fig. 3c, in that the 90th percentile annual mean total phosphorus concentration for these lakes is approximately 24 mg m<sup>-3</sup> (Quinlan et al. 2005).

The form of the relationship between  $J_V$  and Chl *a* or total phosphorus concentration is not clear (Fig. 3b,c).  $J_V$  could reach an upper limit value; if so, it could be in the range 0.227–0.244 g O<sub>2</sub> m<sup>-3</sup> d<sup>-1</sup>. The relationship could be logarithmic, although the number of data points that are >5 mg m<sup>-3</sup> Chl *a* or >15 mg P m<sup>-3</sup> total phosphorus is not sufficient to conclude this; the logarithmic relationship described by Fig. 3b is  $J_V = 0.0657 \ln(\text{Chl } a) - 0.0275$  ( $R^2 = 0.79$ ).

An explanation of the form of the  $J_V$ –trophic state relationship could be based on the volume of water within which the oxygen concentration is reduced as a result of diffusion to the surface of the particles that are responsible for oxygen consumption in the hypolimnetic water; the sphere of water around a particle that is affected in this way is 20 times the radius of the particle (Koch 1971). As the concentration of particles in hypolimnetic water increases as trophic state increases (Baines and Pace 1994), the spheres begin to overlap and so the rate of increase of  $J_V$  decreases (logarithmic relationship) or becomes zero (upper limit value). The effect of particles moving through water during sedimentation on the size of the spheres also needs to be taken into account (Karp-Boss et al. 1996). An upper limit value of 0.23–0.24 g O<sub>2</sub> m<sup>-3</sup> d<sup>-1</sup> is also realistic; use of the oxygen uptake rate for detritus of 1–2 mg O<sub>2</sub> g<sup>-1</sup> hr<sup>-1</sup> from Fig. 1 in Hargrave (1972) and a solids concentration in hypolimnetic water of 5–10 g m<sup>-3</sup> in eutrophic lakes suggests that  $J_V$  would be 0.12–0.48 g O<sub>2</sub> m<sup>-3</sup> d<sup>-1</sup>.

*Areal oxygen consumption rate*— $J_A$  and the two measures of lake trophic state—growing season mean Chl *a* ( $p = 0.93$ , 21 df) and annual mean total phosphorus concentration ( $p = 0.88$ , 23 df)—are not related; the mean  $J_A$  value is  $0.0816 \pm 0.0150$  g O<sub>2</sub> m<sup>-2</sup> whole sediment d<sup>-1</sup> ( $n = 24$ , median = 0.0654; range, 0–0.334; Table 1). These field results suggest that the consumption of oxygen in hypolimnetic sediment does not increase with trophic state, although it should be pointed out that only three lakes have Chl *a* > 7 mg m<sup>-3</sup>, so further results are needed to clarify this indication.

$J_A$  has been widely estimated from sediment cores, and it was found that these values are greater than the field values

by  $\leq 0.73$  g O<sub>2</sub> m<sup>-2</sup> d<sup>-1</sup> (mean = 0.19, median = 0.12) and that this depends on the trophic state of the lake (Fig. 2). The possibility that this difference was due to the assumption in the Livingstone and Imboden (1996) model that vertical diffusion of oxygen can be ignored was evaluated; it was found to change the field  $J_A$  values very little and not enough to account for the lower field  $J_A$  values.

The basis of comparison of the sediment core  $J_A$ , which relates to a part of the sediment at one time during deoxygenation, with the field value, which characterizes the areal oxygen consumption over the hypolimnion for the period of deoxygenation, is not clear, even ignoring the somewhat artificial nature of sediment cores. It is, however, possible to assess how the field value relates to  $J_A$  through a modification of the Livingstone and Imboden (1996) model. It could be that  $J_A$  does not have a constant value in all water strata in the hypolimnion for the whole period of deoxygenation. Its value might be determined by the oxygen consumption rate in the whole sediment,  $R$ , and modified by the oxygen concentration in the water at the sediment surface,  $C(0)$ , which changes during deoxygenation of the stratum. The value of  $C(0)$  is further influenced by oxygen consumption in the water stratum,  $J_V$ , and also by the sediment area to water volume ratio,  $\alpha(Z)$ , in that the higher this value, the greater is the effect of consumption in the sediment on the concentration in the stratum,  $\alpha(Z)J_A$  (Eq. 1). Because  $R$ ,  $J_V$ , and  $\alpha(Z)$  should influence the value of  $J_A$ , and because  $J_A$  changes during deoxygenation, the Livingstone and Imboden model was modified by replacing the constant  $J_A$  parameter with the Hall et al. (1989) model for  $J_A$ . The modified model, described in *Methods*, was used to investigate the effect of  $R$ ,  $J_V$ , and  $\alpha(Z)$  and of the diffusive boundary layer thickness ( $\delta_D$ ), on the average value of  $J_A$  for the whole hypolimnion for the period of deoxygenation.

A low value of  $R$  (20 g O<sub>2</sub> m<sup>-3</sup> whole sediment d<sup>-1</sup>) is needed to produce an average  $J_A$  for the period of deoxygenation between 0.13 and 0.20 g O<sub>2</sub> m<sup>-2</sup> d<sup>-1</sup>, for  $\delta_D = 1$  mm,  $J_V$  over the range 0.02–0.30 g O<sub>2</sub> m<sup>-3</sup> d<sup>-1</sup>, and  $\alpha(Z)$  over the range 0.2–1.0 m<sup>2</sup> m<sup>-3</sup> (Fig. 4; Table 2). The average  $J_A$  decreases as  $J_V$  increases, by 0.007 to 0.058 g O<sub>2</sub> m<sup>-2</sup> d<sup>-1</sup> for a given value of  $\alpha(Z)$ , and, as  $\alpha(Z)$  increases, by 0.003 to 0.041 for a given value of  $J_V$ .

The decrease in average  $J_A$  as  $\alpha(Z)$  increases is not large (0.003–0.041 g O<sub>2</sub> m<sup>-2</sup> d<sup>-1</sup>) compared with the average value (0.13–0.20), but the effect of  $J_A$  not being constant over the hypolimnion does also lead to a further lowering of the field  $J_A$  when fitting a linear regression to the  $J(Z)$ – $\alpha(Z)$  results. This can be shown as follows: The change in average  $J_A$  with  $\alpha(Z)$  for a given  $J_V$  value can be calculated from the results in Table 2; it is  $-0.05$  g O<sub>2</sub> m<sup>-2</sup> sediment d<sup>-1</sup> (m<sup>2</sup> sediment m<sup>-3</sup> water)<sup>-1</sup> for  $J_V \leq 0.10$  g O<sub>2</sub> m<sup>-3</sup> d<sup>-1</sup>, and it can be used to alter the value of  $J_A$  in Eq. 1 as  $\alpha(Z)$  varies from a constant value of 0.20 g O<sub>2</sub> m<sup>-2</sup> d<sup>-1</sup>. In this way, the  $J(Z)$ – $\alpha(Z)$  relationship that would be observed when  $J_A$  is not constant can be reproduced (Fig. 5). The slope of this relationship, ignoring that it is nonlinear, gives the  $J_A$  that would be derived by Eq. 1, the field value, and it is reduced from 0.20 to 0.095 g O<sub>2</sub> m<sup>-2</sup> d<sup>-1</sup>, if the regression uses  $\alpha(Z) \leq 2.0$  m<sup>2</sup> m<sup>-3</sup>, or to

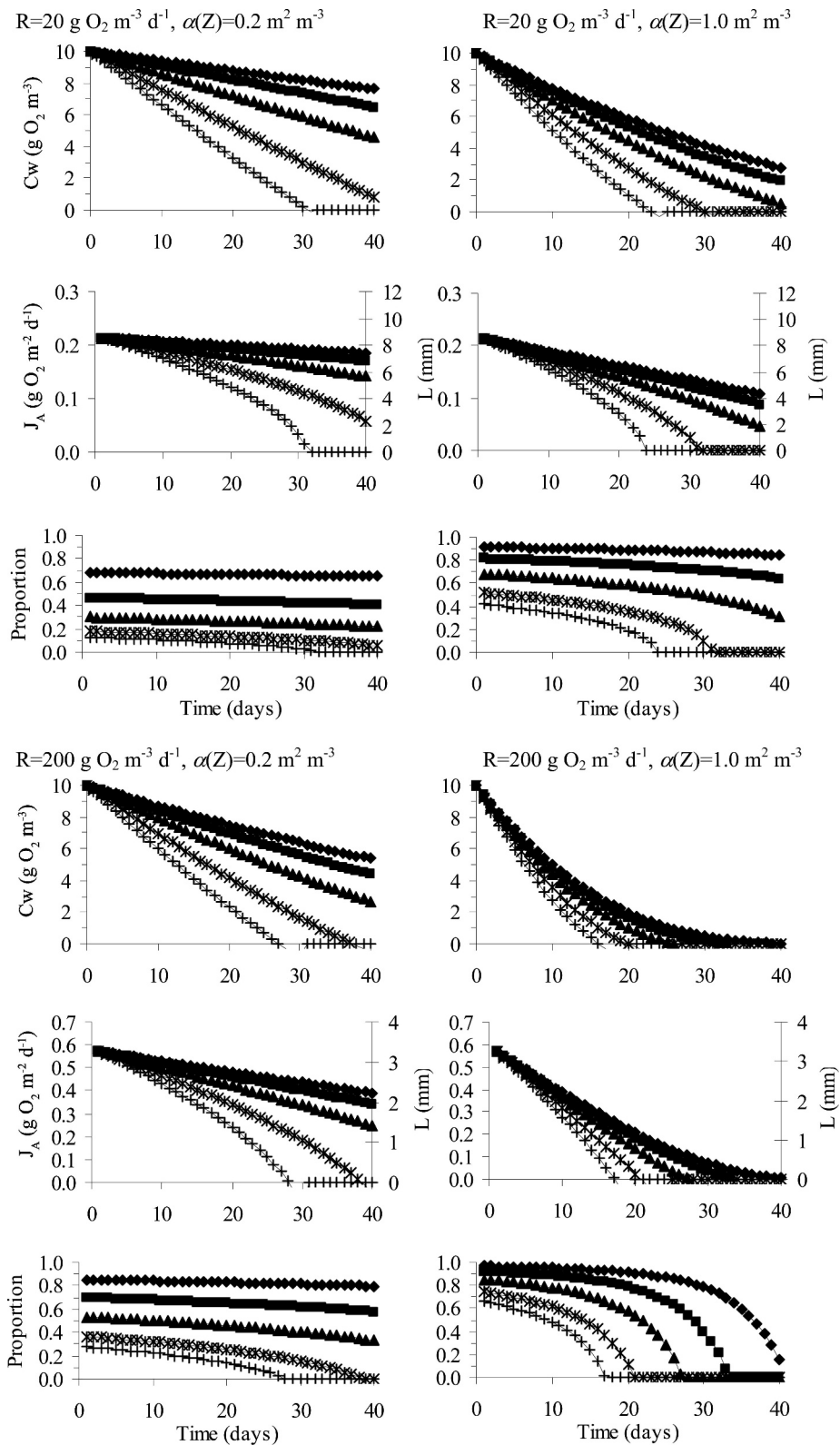


Fig. 4. The variation of oxygen concentration in the water stratum,  $C_w$ ; areal oxygen consumption rate,  $J_A$ ; and consumption of oxygen in the sediment as a proportion of the total consumption in the water stratum with time during deoxygenation as predicted by the modified Livingstone and Imboden (1996) model for five values of the volumetric oxygen consumption rate (0.02, 0.05, 0.10, 0.20, and 0.30  $\text{g O}_2 \text{ m}^{-3} \text{ d}^{-1}$ , corresponding to diamond, square, triangle, X, and +, respectively); two values of the sediment area to water volume ratio,  $\alpha(Z)$ ; two values of the zero-order rate of oxygen consumption in the whole sediment,  $R$ ; and a diffusive boundary layer thickness of 1 mm.



Table 2. The areal oxygen consumption rate ( $J_A$ ,  $\text{g O}_2 \text{ m}^{-2}$  whole sediment  $\text{d}^{-1}$ ) and consumption in the sediment as a proportion of the total consumption in the water stratum, as predicted by the modified Livingstone and Imboden (1996) model for five values of the volumetric oxygen consumption rate ( $J_V$ ,  $\text{g O}_2 \text{ m}^{-3} \text{ d}^{-1}$ ), three values of the sediment area to water volume ratio [ $\alpha(Z)$ ,  $\text{m}^2 \text{ sediment m}^{-3} \text{ water}$ ], two values of the zero-order rate of oxygen consumption in whole sediment ( $R$ ,  $\text{g O}_2 \text{ m}^{-3}$  whole sediment  $\text{d}^{-1}$ ), and two values of the diffusive boundary layer thickness ( $\delta_D$ ).

$J_V$	$J_A$ and consumption in sediment as proportion of total					
$R = 20 \text{ g O}_2 \text{ m}^{-3} \text{ d}^{-1}$ and $\delta_D = 1 \text{ mm}$ for three $\alpha(Z)$						
	0.2	0.4	1.0			
0.02	0.199	0.67	0.190	0.79	0.161	0.89
0.05	0.192	0.43	0.182	0.59	0.152	0.74
0.10	0.179	0.26	0.168	0.40	0.136	0.56
0.20	0.147	0.13	0.136	0.20	0.132	0.37
0.30	0.141	0.08	0.138	0.15	0.135	0.29
$R = 200 \text{ g O}_2 \text{ m}^{-3} \text{ d}^{-1}$ and $\delta_D = 1 \text{ mm}$ for three $\alpha(Z)$						
	0.2	0.4	1.0			
0.02	0.480	0.83	0.406	0.89	0.230	0.82
0.05	0.457	0.64	0.381	0.74	0.262	0.73
0.10	0.415	0.45	0.336	0.55	0.284	0.75
0.20	0.338	0.24	0.327	0.36	0.317	0.55
0.30	0.343	0.17	0.335	0.28	0.322	0.45
$R = 20 \text{ g O}_2 \text{ m}^{-3} \text{ d}^{-1}$ and $\delta_D = 5 \text{ mm}$ for three $\alpha(Z)$						
	0.2	0.4	1.0			
0.02	0.147	0.59	0.140	0.74	0.122	0.86
0.05	0.140	0.36	0.134	0.52	0.115	0.69
0.10	0.129	0.20	0.122	0.33	0.103	0.49
0.20	0.102	0.09	0.095	0.15	0.089	0.28
0.30	0.094	0.05	0.094	0.10	0.091	0.22
$R = 200 \text{ g O}_2 \text{ m}^{-3} \text{ d}^{-1}$ and $\delta_D = 5 \text{ mm}$ for three $\alpha(Z)$						
	0.2	0.4	1.0			
0.02	0.228	0.69	0.208	0.80	0.159	0.88
0.05	0.215	0.46	0.192	0.60	0.148	0.72
0.10	0.192	0.28	0.174	0.40	0.129	0.52
0.20	0.163	0.12	0.131	0.19	0.122	0.33
0.30	0.138	0.08	0.134	0.14	0.129	0.27

0.145 if  $\leq 1.0 \text{ m}^2 \text{ m}^{-3}$  is used. This is a considerable reduction in the average value of  $J_A$  estimated from field results, especially if the larger  $\alpha(Z)$  range is used, and it could be the main reason why the field  $J_A$  is less than the sediment cores value if  $R$  is low and it is assumed that the sediment core value is an estimate of  $J_A$  at some time during the earlier part of the deoxygenation period (Fig. 4). The nonlinear  $J(Z)-\alpha(Z)$  behavior when the  $\alpha(Z)$  range is large might explain why the strong linear increase of  $J(Z)$  with  $\alpha(Z)$  in five lakes in Quebec and one in central Ontario did not continue to the highest  $\alpha(Z)$  value (Lovering, Stukely, Lyster, Orford, North Memphremagog, and Solitaire; Figs. 1, 5).

The value of  $J_A$  is not constant but decreases during deoxygenation (Fig. 4), especially during the latter stages, and the proportion of the total oxygen consumption that occurs in the sediment is also not constant, becoming smaller as deoxygenation progresses, particularly at high  $J_V$  and low  $\alpha(Z)$  (Fig. 4; Table 2).

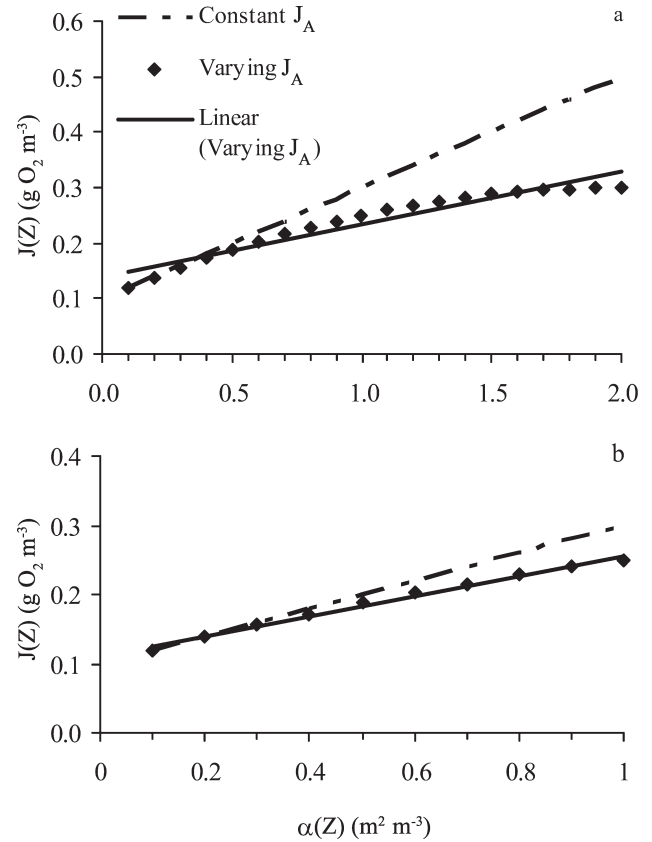


Fig. 5. The variation of oxygen depletion rate,  $J(Z)$ , at water depth  $Z$  with the sediment area to water volume ratio,  $\alpha(Z)$ , for a constant value of the areal oxygen consumption rate,  $J_A$  ( $0.20 \text{ g O}_2 \text{ m}^{-2} \text{ d}^{-1}$ ) and for  $J_A$  varying with  $\alpha(Z)$  at a rate of  $-0.05 \text{ g O}_2 \text{ m}^{-2} \text{ sediment m}^{-3} \text{ water}^{-1}$ . The Livingstone and Imboden (1996) model was used to calculate  $J(Z)$  with the varying  $J_A$  values and a volumetric oxygen consumption rate of  $0.10 \text{ g O}_2 \text{ m}^{-3} \text{ d}^{-1}$ . The linear regression between  $J(Z)$  and  $\alpha(Z)$  that would be obtained with varying  $J_A$ , but assuming a linear relationship, is also shown. (a) The behavior if an  $\alpha(Z)$  range of  $2.0 \text{ m}^2 \text{ m}^{-3}$  is used and (b) if  $1.0 \text{ m}^2 \text{ m}^{-3}$  is used.

The reduction in the average  $J_A$  at higher values of  $J_V$  and  $\alpha(Z)$  is greater when  $R$  is higher ( $200 \text{ g O}_2 \text{ m}^{-3}$  whole sediment  $\text{d}^{-1}$ ); the average  $J_A$  decreases by  $-0.092$  to  $0.142 \text{ g O}_2 \text{ m}^{-2} \text{ d}^{-1}$  at the higher values of  $J_V$  for a given value of  $\alpha(Z)$  and by  $0.008$  to  $0.250$  at the higher values of  $\alpha(Z)$  for a given value of  $J_V$  (Fig. 4; Table 2). This could be one reason for the low field  $J_A$  compared with the sediment core value if  $R$  is high and the sediment core value applies to a point earlier in the deoxygenation. The effect of  $J_A$  not being constant as  $\alpha(Z)$  increases is also greater when  $R$  is larger (Table 2); the slope of the  $J_A-\alpha(Z)$  relationship for a given value of  $J_V$  is around  $-0.20 \text{ g O}_2 \text{ m}^{-2} \text{ sediment m}^{-3} \text{ water}^{-1}$  for  $J_V \leq 0.10 \text{ g O}_2 \text{ m}^{-3} \text{ d}^{-1}$ , and so the linear  $J(Z)-\alpha(Z)$  relationship when  $J_A$  is not constant leads to a reduced estimate of the average  $J_A$  from  $0.48$  to  $0.060 \text{ g O}_2 \text{ m}^{-2} \text{ d}^{-1}$  if the regression uses  $\alpha(Z) \leq 2.0 \text{ m}^2 \text{ m}^{-3}$  or to  $0.260$  if  $\alpha(Z) \leq 1.0 \text{ m}^2 \text{ m}^{-3}$  is used. The behavior is similar to that in Fig. 5, but the reduction in average  $J_A$  is greater with the higher  $R$  values. This could be a second reason for the lower field  $J_A$  values if  $R$  is high.

Table 3. The zero-order rate of oxygen consumption in whole sediment ( $R$ ,  $\text{g O}_2 \text{ m}^{-3} \text{ d}^{-1}$ ) in lakes available in the literature or derived by Eq. 5 from results in the literature, along with the observed values of the depth of oxygen penetration in the sediment ( $L$ ), oxygen concentration at the sediment-water interface [ $C(0)$ ] and the ratio of  $C(0)$  to the oxygen concentration in the water stratum above the diffusive boundary layer ( $C_w$ ). The method used and the spatial resolution of the measurements are also given.

Reference	Lake	Site	Method	Resolution ( $\mu\text{m}$ )	Date	$R$ ( $\text{g O}_2$ $\text{m}^{-3} \text{ d}^{-1}$ )	$L$ (mm)	$C(0)$ ( $\text{g O}_2 \text{ m}^{-3}$ )	$C(0):C_w$
Muller et al. (2002)	Alpnach	Profundal	in situ	50	24 Mar 98	99	3.7	9.9	0.89
				50	21 Apr 98	147	2.9	9.8	0.86
				50	27 May 98	384	1.7	7.3	0.81
				50	14 Jul 98	458	1.3	5.4	0.74
					25 Aug 98	205	1.3	2.3	0.79
Lorke et al. (2003)	Alpnach	Profundal	in situ	50	14–15 Aug 02	168	0.41		
Sweets et al. (1986)	L302S (fig. 2)	Littoral	core	200	Summer 83	16	12.6	7.7	0.90
Sweets and Cappenberg (1988)	Kievitsbuurt (fig. 1)	Littoral	core	200	1986	329	3.2	9.6	0.79
	L305 (fig. 2)	Littoral	core	250	Jun 86	283	2.25	6.9	0.90
	L302S (fig. 3)	Littoral	in situ	250	Jun 86	152	4.5	11.2	1.21
Danielopol and Niederreiter (1990)	Mondsee (fig. 6)	Profundal, 20 m	core	1000	Apr 88	249	5	16.5	0.92
	Mondsee (fig. 7)	Profundal, 6 m	core	1000	Feb 88	22	10	10.0	0.63
Martin et al. (1993)	Baikal (fig. 2)	Profundal, 22 m	core	500	Aug 91	47	5.5	5.4	
Huttunen et al. (2006)	Soiviojarvi	Profundal, 9.8 m	core	1000	Aug 94	23	6		
	Postilampi	Profundal, 32.5 m	core	1000	Aug 95	13	5		
	Lokka	Profundal, 8.8 m	core	1000	Aug 95	6.6	4		
	Tuusulanjarvi	Profundal, 7.8 m	core	1000	Aug 95	14	3		

Note.  $R$  was estimated by Müller et al. (2002) by fitting in situ oxygen microprofile results for the profundal of Lake Alpnach to Eq. 5. The mean in situ value for a 25-h period during 14–15 August 2002 in the same lake was measured by Lorke et al. (2003) and B. Müller (pers. comm.). Further values for  $R$  were derived by fitting Eq. 5 to published microprofile results and adjusting the value of  $R$  to minimize the variance between observed and predicted concentrations. No porosity values were given, so a value of 0.98 was assumed for the sediment surface between 0.4 and 1.5 mm depth.

An example would be the low field  $J_A$  values in the three eutrophic Irish lakes, in that the maximum  $\alpha(Z)$  value is  $>1.0 \text{ m}^2 \text{ m}^{-3}$  ( $1.40\text{--}3.40 \text{ m}^2 \text{ m}^{-3}$ ; Table 1).

An increase in  $\delta_D$  to 5 mm when  $R$  is high causes a further and considerable decrease in the mean value of  $J_A$  and, assuming  $J_A$  does not vary with  $\alpha(Z)$ , reduces the value even further (results not shown). The effect of this higher  $\delta_D$  value on reducing the average  $J_A$  value is less when  $R$  is low.

The use of the modified model, in which  $J_A$  is not constant with depth and time, has shown that consumption of oxygen in a water stratum, as represented by  $J_V$ , and consumption in the sediment, as represented by  $\alpha(Z)J_A$ , causes consumption in the sediment to decrease as  $J_V$  and  $\alpha(Z)$  increase. The effect of consumption is to reduce  $C(0)$ , and as  $C(0)$  decreases, so does  $J_A$  (Eq. 2).  $J_A$  should be proportional to  $C(0)^{0.5}$ , and this has been observed in sediment cores by Provini (1975). Furthermore, the effect of  $J_A$  decreasing with  $\alpha(Z)$  also leads to a lower average  $J_A$  value derived from a linear relationship between  $J(Z)$  and  $\alpha(Z)$ . These two reasons are likely to be responsible for the field  $J_A$  values being lower than those estimated from sediment cores; it could be that the second is more important if  $R$  is low and that both are involved if  $R$  is high. Although further analysis of the extrapolation of a sediment core value to the whole period of deoxygenation is needed, at this stage, the field  $J_A$  value appears more relevant in the context of modeling deoxygenation in lake hypolimnia. It integrates consumption over the sediment in the hypolimnion and over the period of deoxygenation. Single field values for  $J_A$  and  $J_V$  in a lake and the condensed linear model described by Eq. 1 do generally reproduce the observed variation of  $J(Z)$  with  $\alpha(Z)$  in the 32 lakes well (Fig. 1). We have identified the limitations of the model and they are not due to the difference between the field and sediment core  $J_A$  values but to the occasional bias estimating  $J(V)$  with nonlinear  $J(V)\text{--}\alpha(Z)$  behavior in lakes with a high  $\alpha(Z)$  range.

*Oxygen consumption rate in the whole sediment*—Because the oxygen consumption rate in the whole sediment ( $R$ ) is an important variable in the modified Livingstone and Imboden (1996) model, it is useful to know its value in lake sediment (Table 3).  $R$  varies from 7 to  $460 \text{ g O}_2 \text{ m}^{-3}$  whole sediment  $\text{d}^{-1}$ , and a histogram of the 17 values from the 10 lakes shows that they are generally around 20 or  $200\text{--}300 \text{ g O}_2 \text{ m}^{-3}$  whole sediment  $\text{d}^{-1}$ ; 20 and 200 were used in the application of the modified model. The depth of oxygen penetration in the sediment ( $L$ ) varies from 0.41 to 12.6 mm (mean = 4.3, median = 3.7). With 11 values of the  $C(0):C_w$  ratio, the mean = 0.82, with all values  $>0.62$ . As is expected from Eq. 3, the relationship is strong between  $L$  and  $[C(0)/R]^{0.5}$  ( $R^2 = 0.86$ ,  $p < 0.001$ , 14 df), whereas that between  $L$  and  $R^{0.5}$  is much weaker ( $R^2 = 0.25$ ,  $p = 0.043$ , 16 df).

There is some evidence that the value of  $R$  changes seasonally. It increased from 99 to  $458 \text{ g O}_2 \text{ m}^{-3} \text{ d}^{-1}$  from March to July 1998 in the profundal of Lake Alpnach and then decreased; the higher values lasted for approximately 2 months (Table 3). The evidence is stronger for seasonal changes in  $R$  in coastal sediment. Glud et al. (2003) showed directly that  $R$  varied seasonally in the sediment of Aarhus Bay during 1990–

1992, from a low of 80 to highs of 480 and  $960 \text{ g O}_2 \text{ m}^{-3} \text{ d}^{-1}$ , during periods of intensified sedimentation; the higher values lasted 2 to 3 months. This evidence suggests that  $R$  does vary seasonally because of the sedimentation of fresh organic matter and variations in sediment temperature. The effect of fresh organic matter in the sediment resulted in higher  $R$  values for around 2 months in Lake Alpnach and 2 to 3 months in Aarhus Bay. Because  $R$  seems to vary, it might be as difficult as it was with the sediment core  $J_A$  to characterize the consumption rate of oxygen in sediment for the period of deoxygenation with a single value.

The use of the Livingstone and Imboden (1996) model has produced field values for the volumetric oxygen consumption rate ( $J_V$ ) and areal oxygen consumption rate ( $J_A$ ) that apply to the deoxygenation period in the hypolimnia of 32 lakes. Analysis of a modified model showed that  $J_A$  is not constant with depth and time in the hypolimnion, and it was suggested that the field  $J_A$  value integrates the areal consumption over the hypolimnion and over the deoxygenation period. If it is taken that the field  $J_A$  value integrates oxygen consumption in the hypolimnion and that  $J(Z)$  varies linearly with  $\alpha(Z)$ , then the Livingstone and Imboden model is a general condensed model for oxygen depletion in lake hypolimnia. It can be used to predict the changing oxygen concentration–water depth profiles during deoxygenation in thermally stratified lakes with the use of a constant value for  $J_A$  ( $0.0816 \pm 0.0150 \text{ g O}_2 \text{ m}^{-2}$  whole sediment  $\text{d}^{-1}$ ) and the relationship between  $J_V$  and lake trophic state (Fig. 3b,c) to predict how  $J(Z)$  varies with  $\alpha(Z)$  in a lake according to Eq. 1. There is some empiricism in this approach, in the use of the relationships in Fig. 3, and so it is similar in some ways to the model of Molot et al. (1992).

#### Acknowledgments

We thank Peter Devlin, Colin McQuillan, and Mark Poole for their assistance with the fieldwork. We also thank two anonymous reviewers for their comments and suggestions, which improved the manuscript greatly. The Natural Environment Research Council supported this work through grant NE/C514315/1.

#### References

- BAINES, S. B., AND M. L. PACE. 1994. Relationships between suspended particulate matter and sinking flux along a trophic gradient and implications for the fate of plankton primary production. *Can. J. Fish. Aquat. Sci.* **51**: 25–36.
- BEUTEL, M. W. 2003. Hypolimnetic anoxic and sediment oxygen demand in California drinking water reservoirs. *Lake Reservoir Manag.* **19**: 208–221.
- BOULDIN, D. R. 1968. Models for describing the diffusion of oxygen and other mobile constituents across the mud–water interface. *J. Ecol.* **56**: 77–87.
- BROECKER, W. S., AND Y.-H. PENG. 1974. Gas exchange rates between air and sea. *Tellus* **26**: 21–35.
- CLARKE, B. J., P. J. DILLON, L. A. MOLOT, AND H. E. EVANS. 2002. Application of a hypolimnetic profile model to lakes in Ontario. *Lake Reservoir Manag.* **18**: 32–43.
- COHEN, J. 1988. *Statistical power analysis for the behavioural sciences*. Lawrence Erlbaum.
- CORNETT, R. J. 1989. Predicting changes in hypolimnetic oxygen concentrations with phosphorus retention, temperature and morphometry. *Limnol. Oceanogr.* **34**: 1359–1366.

- , AND F. H. RIGLER. 1980. The areal hypolimnetic oxygen deficit: An empirical test of the model. *Limnol. Oceanogr.* **25**: 672–679.
- , AND ———. 1987. Vertical transport of oxygen into the hypolimnion of lakes. *Can. J. Fish. Aquat. Sci.* **44**: 852–858.
- DANIELOPOL, D. L., AND R. NIEDERREITER. 1990. The measurement of dissolved oxygen in lacustrine sediments. Laboratory set-up and techniques. *Bull. Inst. Geol. Bassin d'Aquitaine, Bordeaux* **47**: 287–296.
- DAVISON, W. 1993. Iron and manganese in lakes. *Earth Sci. Rev.* **34**: 119–163.
- DINSMORE, W. P., AND E. E. PREPAS. 1997. Impact of hypolimnetic oxygenation on profundal macroinvertebrates in a eutrophic lake in Central Alberta. II. Changes in *Chironomus* spp. abundance and biomass. *Can. J. Fish. Aquat. Sci.* **54**: 2170–2181.
- Directive 2000/60/EC of the European Parliament and of the Council of 23 October 2000 establishing a framework for community action in the field of water policy. *Off. J. Eur. Communities* **L327**: 1–72.
- EVANS, D. O., K. H. NICHOLLS, Y. C. ALLEN, AND M. J. MCMURTRY. 1996. Historical land use, phosphorus loading, and loss of fish habitat in Lake Simcoe, Canada. *Can. J. Fish. Aquat. Sci.* (suppl. 1) **53**: 194–218.
- GLUD, R. N., J. K. GUNDERSEN, H. ROY, AND B. B. JORGENSEN. 2003. Seasonal dynamics of benthic O<sub>2</sub> uptake in a semienclosed bay: Importance of diffusion and faunal activity. *Limnol. Oceanogr.* **48**: 1265–1276.
- GRANELI, W. 1977. Measurement of sediment oxygen uptake in the laboratory using undisturbed sediment cores. *Vatten* **3**: 251–265.
- HALL, P. O. J., L. G. ANDERSON, M. M. RUTGERS VAN DER LOEFF, B. SUNDBY, AND S. F. G. WESTERLUND. 1989. Oxygen uptake kinetics in the benthic boundary layer. *Limnol. Oceanogr.* **34**: 734–746.
- HARGRAVE, B. T. 1972. Aerobic decomposition of sediment and detritus as a function of particle surface area and organic content. *Limnol. Oceanogr.* **17**: 583–596.
- HUTCHINSON, G. E. 1938. On the relation between the oxygen deficit and the productivity and the typology of lakes. *Int. Rev. Ges. Hydrobiol.* **36**: 336–355.
- . 1957. *A treatise on limnology, v. 1. Geography, physics and chemistry*, Wiley.
- HUTTUNEN, J. T., T. S. VAISANEN, S. K. HELLSTEN, AND P. J. MARTIKAINEN. 2006. Methane fluxes at the sediment–water interface in some boreal lakes and reservoirs. *Boreal Environ. Res.* **11**: 27–34.
- JASSBY, A., AND T. POWELL. 1975. Vertical patterns of eddy diffusion during stratification in Castle Lake, California. *Limnol. Oceanogr.* **20**: 530–543.
- KARP-BOSS, L., E. BOSS, AND P. A. JUMARS. 1996. Nutrient fluxes to planktonic osmotrophs in the presence of fluid motion. *Oceanogr. Mar. Biol. Annu. Rev.* **34**: 71–107.
- KOCH, A. L. 1971. The adaptive responses of *Escherichia coli* to a feast and famine existence. *Adv. Microb. Physiol.* **6**: 147–217.
- LIVINGSTONE, D. M., AND D. M. IMBODEN. 1996. The prediction of hypolimnetic oxygen profiles: A plea for a deductive approach. *Can. J. Fish. Aquat. Sci.* **53**: 924–932.
- LORKE, A., B. MULLER, M. MAERKI, AND A. WUEST. 2003. Breathing sediments: The control of diffusive transport across the sediment–water interface by periodic boundary-layer turbulence. *Limnol. Oceanogr.* **48**: 2077–2085.
- MAERKI, M., B. WEHRLI, C. DINKEL, AND B. MULLER. 2004. The influence of tortuosity on molecular diffusion of freshwater sediments of high porosity. *Geochim. Cosmochim. Acta* **68**: 1519–1528.
- MARKER, A. F. H., C. A. CROWTHER, AND R. J. M. GUNN. 1980. Methanol and acetone as solvents for estimating chlorophyll *a* and phaeopigments by spectrophotometry. *Arch. Hydrobiol. Beih. Ergebn. Limnol.* **14**: 52–69.
- MARTIN, P., B. GODDEERIS, AND K. MARTENS. 1993. Oxygen concentration profiles in soft sediment of Lake Baikal (Russia) near the Selenga delta. *Freshw. Biol.* **29**: 343–349.
- MATHIAS, J. A., AND J. BARCIA. 1980. Factors controlling oxygen depletion in ice-covered lakes. *Can. J. Fish. Aquat. Sci.* **37**: 185–194.
- MOLOT, L. A., P. J. DILLON, B. J. CLARKE, AND B. P. NEARY. 1992. Predicting end-of-summer oxygen profiles in stratified lakes. *Can. J. Fish. Aquat. Sci.* **49**: 2363–2372.
- MORTIMER, C. H. 1942. The exchange of dissolved substances between mud and water in lakes. III and IV, summary and references. *J. Ecol.* **30**: 147–201.
- MÜLLER, B., M. MARKI, C. DINKEL, R. STIERLI, AND B. WEHRLI. 2002. In situ measurement on lake sediment using ion-selective electrodes with a profiling lander system, p. 126–143. *In* M. Tallefert and T. F. Rozan [eds.], *Environmental electrochemistry*. American Chemical Society.
- NURNBERG, G. 1995. Quantifying anoxia in lakes. *Limnol. Oceanogr.* **40**: 1100–1111.
- NURNBERG, G. K. 1984. The prediction of internal phosphorus load in lakes with anoxic hypolimnia. *Limnol. Oceanogr.* **29**: 111–124.
- PACE, M. L., AND Y. T. PRAIRIE. 2005. Respiration in lakes, p. 103–121. *In* P. A. del Giorgio and P. J. le B Williams [eds.], *Respiration in aquatic ecosystems*. Oxford University Press.
- PROVINI, A. 1975. Sediment respiration in six Italian lakes in different trophic conditions. *Verh. Int. Ver. Limnol. Theor. Angew.* **19**: 1313–1318.
- QUAY, P. D., W. S. BROECKER, R. H. HESSLEIN, AND D. W. SCHINDLER. 1980. Vertical diffusion rates determined by tritium tracer experiments in the thermocline and hypolimnion of two lakes. *Limnol. Oceanogr.* **25**: 201–218.
- QUINLAN, R., A. M. PATERSON, J. P. SMOL, M. S. V. DOUGLAS, AND B. J. CLARK. 2005. Comparing different methods of calculating volume-weighted hypolimnetic oxygen (VWHO) in lakes. *Aquat. Sci.* **67**: 97–103.
- STEFAN, H. G., AND X. FANG. 1994. Dissolved oxygen model for regional lake analysis. *Ecol. Model.* **71**: 37–68.
- STROM, K. M. 1931. Feforvatn. A physiographic and biological study of a mountain lake. *Arch. Hydrobiol.* **22**: 491–536.
- SWEERTS, J. P., J. W. M. RUDD, AND C. A. KELLY. 1986. Metabolic activities in flocculent surface sediments and underlying sandy littoral sediments. *Limnol. Oceanogr.* **31**: 330–338.
- SWEERTS, J.-P. R. A., AND T. E. CAPPENBERG. 1988. Use of microelectrodes for measuring oxygen profiles in lake sediments. *Arch. Hydrobiol. Beih. Ergebn. Limnol.* **31**: 365–371.
- , V. ST. LOUIS, AND T. E. CAPPENBERG. 1989. Oxygen concentration profiles and exchange in sediment cores with circulated overlying water. *Freshw. Biol.* **21**: 401–409.
- VINCENT, W. F., M. M. GIBBS, AND S. J. DRYDEN. 1984. Accelerated eutrophication in a New Zealand lake: Lake Rotoiti, Central North Island. *N. Z. J. Mar. Freshw. Res.* **18**: 431–440.
- VOLLENWEIDER, R. A., AND L. L. JANUS. 1982. Statistical models for predicting hypolimnetic oxygen depletion rates. *Mem. Ist. Ital. Idrobiol.* **40**: 1–24.
- WELCH, E. B., AND M. A. PERKINS. 1979. Oxygen deficit–phosphorus loading relation in lakes. *J. Water Pollut. Control Fed.* **51**: 2823–2828.

Associate editor: Bo Thamdrup

Received: 11 March 2008

Accepted: 21 November 2008

Amended: 08 February 2009